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# Removal of sulfur dioxide by fly ash/CaO/CaSO4 sorbents

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#### **Abstract**

This study presents findings from an experimental investigation of the influences of several factors on the desulfurization activity of sorbents synthesized from coal fly ash, CaO, and CaSO<sub>4</sub>. Specifically, the effects of specific surface area  $(16.1-133.3 \text{ m}^2/\text{g})$ , reaction temperature (60–300 °C), feed concentration of  $SO_2$  (500–2500 ppm), and feed concentration of NO (0–750 ppm) were elucidated using a fixed bed reactor. In general, the sorbent desulfurization activity increased with increasing specific surface area, reaction temperature, and NO concentration but with decreasing  $SO<sub>2</sub>$  concentration. The physical and chemical properties of sorbent before and after  $SO<sub>2</sub>$  capture were characterized through scanning electron microscopy, X-ray diffraction, and Fourier transform infrared analyses. © 2005 Elsevier B.V. All rights reserved.

*Keywords:* Desulfurization; Fixed bed reactor; Flue gas; Fly ash; Sorbent; Sulfur dioxide

## **1. Introduction**

Sulfur in coal is converted to sulfur dioxide  $(SO<sub>2</sub>)$  during combustion.  $SO_2$  is a precursor to acid rain which is one of the most widespread forms of pollution worldwide. In addition,  $SO_2$ causes the formation of secondary particles in the atmosphere that impair visibility. To minimize the adverse impacts of  $SO<sub>2</sub>$ emissions on the environment, much effort has been put into developing flue gas desulfurization (FGD) technologies over the past few decades for the control and abatement of  $SO<sub>2</sub>$  emissions by coal-fired power plants and boilers of industrial processes. For example, dry FGD processes use calcium-based sorbents, usually hydrated lime  $(Ca(OH)_2)$ , as a reagent to remove  $SO_2$ from flue gas.

Recent research in this area has focused on synthesizing different types of calcium-based sorbent from mixtures of coal fly ash/Ca(OH) $_2$ /CaO/CaSO<sub>4</sub>. The use of coal fly ash as the base material in sorbent synthesis is attractive both economically and environmentally, as it is the most voluminous by-product of coal-fired power plants. A number of studies have shown that mixing coal fly ash with Ca(OH)<sub>2</sub> or CaO in a hydration process produces sorbents with SO<sub>2</sub> capture capacity higher than

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that of hydrated lime [\[1–11\]. I](#page-5-0)n addition, it has been shown that addition of  $CaSO<sub>4</sub>$  to the preparation mixture has a positive or negative effect on the sorbent desulfurization activity, depending on the specific hydration conditions being used [\[12–14\].](#page-6-0) These coal fly ash-based sorbents have been successfully used in a dry-type FGD system to treat  $644\,000 \,\mathrm{m}^3/\mathrm{h}$  of flue gas [\[4\].](#page-5-0)

The extent of  $SO<sub>2</sub>$  capture varies significantly for different types of sorbent synthesized from coal fly ash/Ca(OH)2/CaO/ CaSO4. The structural properties of these sorbents (e.g., specific surface area and porosity) are believed to play an important role in  $SO<sub>2</sub>$  capture. In general, it has been shown that higher desulfurization activity correlates with higher specific surface area [\[2,11\],](#page-5-0) although some studies have produced inconclusive results [\[8,9\].](#page-6-0) These previous results suggest that further study on the role of specific surface area in  $SO<sub>2</sub>$  capture is warranted. In this work, we report an experimental investigation of  $SO<sub>2</sub>$ removal by sorbents having a range of specific surface areas prepared from coal fly ash, CaO, and CaSO4. The other key areas of interest are the effects of reactor operating conditions such as temperature and feed concentrations of  $SO<sub>2</sub>$  and NO on the extent of SO<sub>2</sub> removal and characterization of sorbent before and after  $SO<sub>2</sub>$  capture through scanning electron microscopy (SEM), X-ray diffraction (XRD), and Fourier transform infrared (FTIR) analyses to gain insight into the mechanism of  $SO<sub>2</sub>$ removal.

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<span id="page-1-0"></span>



# **2. Materials and methods**

## *2.1. Sorbent preparation*

The raw materials used to prepare the sorbents were coal fly ash, CaO, and CaSO4. The coal fly ash was provided by a coal-fired power plant of Tenaga Nasional Berhad, Malaysia. The chemical composition of the fly ash was  $60\%$  SiO<sub>2</sub>, 20% Al<sub>2</sub>O<sub>3</sub>, 4.7% Fe<sub>2</sub>O<sub>3</sub>, 3.0% CaO, 1.1% K<sub>2</sub>O, 1.0% MgO, 7.5% C, 2.4% others, and 0.3% ignition loss. The CaO and  $CaSO<sub>4</sub>$  were obtained from BDH Laboratories, England. The BET surface area of these raw materials was analyzed using an Autosorb 1C Quantachrome analyzer and the results are listed in Table 1.

To evaluate the effect of specific surface area on  $SO<sub>2</sub>$  capture, sorbents having a range of specific surface area were prepared by varying some of the preparation conditions. Specifically, a full factorial design was used to synthesize sorbents of different surface areas by varying three experimental variables: hydration period, amount of coal fly ash, and amount of CaSO4. The amount of CaO used was kept constant. Table 2 displays the matrix for the  $3<sup>3</sup>$  factorial design. The procedure used to prepare the sorbents comprised the following steps. A fixed amount of CaO (5 g) was added to 100 ml of water at 65 ◦C. Different amounts of coal fly ash and CaSO4 were added to the slurry simultaneously, according to the design matrix shown in Table 2. The slurry was then maintained at  $95^{\circ}$ C for a period of time for the hydration process to occur. The hydration period for different runs was varied according to the design matrix shown in Table 2. The resulting slurry was filtered and dried at 200 ◦C for 2 h. The BET surface area of the resulting sorbents was analyzed as described above and the results are tabulated in Table 2.

To study the effects of reactor operating conditions such as temperature,  $SO<sub>2</sub>$  concentration, and NO concentration on the extent of  $SO<sub>2</sub>$  capture, a reference sorbent having a fixed specific surface area was prepared. The reference sorbent was synthesized by mixing  $5 g CaO$ ,  $13.7 g flyash$ , and  $7.4 g CaSO<sub>4</sub>$  for a hydration period of 10 h. The resulting sorbent had a specific surface area of  $64.5 \text{ m}^2/\text{g}$ .

#### *2.2. Desulfurization experiments*

Desulfurization experiments were performed using a fixed bed test rig, as shown in Fig. 1. The reaction zone was contained in a 0.13 m diameter stainless steel tube fitted in a furnace for isothermal operation. The sorbent  $(0.7 g)$  was packed in the center of the reactor supported by glass wool. A  $N_2$  stream passed through a humidification system consisting of two 250 ml conical flasks immersed in a water bath at constant temperature. After humidification, the  $N_2$  stream was mixed with a gas stream con-



sisting of  $SO_2$ , NO,  $O_2$ , and  $CO_2$ . The combined feed gas flowed through the reactor at a flow rate of  $150$  ml/min. The  $SO<sub>2</sub>$  concentration of the effluent gas was continuously monitored by an Enerac 2000E flue gas analyzer. Two or three replicate measurements were made for each sorbent activity test and the relative standard deviation was found to be less than 3%. For clarity, only the average values are presented in this paper. The desulfurization activity is expressed in three different ways: maximum utilization of CaO (mg captured  $SO_2/g$  CaO) in the sorbent when



Fig. 1. Schematic diagram showing the experimental set-up for fixed bed desulfurization.



it is completely exhausted (i.e., percent removal of  $SO_2 = 0$ )[\[13\];](#page-6-0) duration the sorbent could maintain above  $80\%$  removal of  $SO<sub>2</sub>$ per gram CaO (min/g CaO) [\[5\];](#page-5-0) percent removal of  $SO_2$ .

To evaluate the effect of specific surface area on  $SO_2$  capture, a range of sorbents having different surface areas were selected from [Table 2](#page-1-0) for the activity test. The feed gas composition at  $100\degree$ C was maintained at 1000 ppm SO<sub>2</sub>, 500 ppm NO, 5% O<sub>2</sub>, 12%  $CO<sub>2</sub>$ , and the balance N<sub>2</sub>. To study the effects of reactor operating conditions on  $SO_2$  capture, the fixed bed reactor packed with the reference sorbent was subjected to a feed gas consisting of 500–2500 ppm  $SO_2$ , 0–750 ppm NO, 5%  $O_2$ , 12%  $CO<sub>2</sub>$ , and the balance N<sub>2</sub>. The range of reaction temperature tested was 60–300 °C. Additional desulfurization experiments were carried out at 150 °C using the fixed bed reactor packed with each of the raw materials (coal fly ash, CaO, or  $CaSO<sub>4</sub>$ ) and a feed gas consisting of  $1000$  ppm  $SO<sub>2</sub>$ ,  $500$  ppm NO,  $5\%$  $O_2$ , 12%  $CO_2$ , and the balance  $N_2$ .

#### *2.3. Physical and chemical analyses*

The specific surface area and pore volume distribution of sorbents were determined using the BET and BJH methods, respectively, on an Autosorb 1C Quantachrome analyzer. A Leica Cambridge S360 scanning electron microscope with 15 kV of accelerating voltage was used to examine the sorbent surface morphologies. The CaO content of sorbent was determined using a Rigaku RIX 3000 X-ray spectrometer. XRD analysis was carried out using a Philips PW 1820 X-ray diffractometer with Cu K $\alpha$  radiation in the diffraction angle (2θ) range of 5–90 $\degree$  at a sweep rate of 3°/min. FTIR spectra were recorded on a Perkin-Elmer FTIR 2000 spectrophotometer. The amount of sample used in the XRD and FTIR analyses was kept constant so that the peaks obtained in the spectra can be used to represent the relative concentrations of a chemical species.

## **3. Results and discussion**

#### *3.1. Effect of specific surface area*

A number of previous studies have produced inconclusive results with regard to the influence of specific surface area on the desulfurization activity of sorbents derived from fly ash/Ca(OH) $_2$ /CaO/CaSO<sub>4</sub>. Some studies reported that desulfurization activity increased with increasing specific surface area [\[2,11\]](#page-5-0) while others did not find a direct relationship between these two parameters [\[8,9\].](#page-6-0) To elucidate the role of specific surface area in  $SO<sub>2</sub>$  capture, sorbents having surface areas in the range of  $16.1-133.3 \text{ m}^2/\text{g}$  were synthesized from coal fly ash/CaO/CaSO4 by varying some of the preparation conditions according to a full factorial design, as shown in [Table 2.](#page-1-0) Selected sorbents were tested for their desulfurization activity which is defined as the maximum utilization of CaO (mg captured  $SO_2/g$  CaO) in the sorbent when it is completely exhausted (i.e., percent removal of  $SO_2 = 0$ ). Fig. 2 shows the desulfurization activity of sorbents versus their specific surface area. The desulfurization activity is seen to increase exponentially from 53.6 to 244.7 mg captured  $SO_2/g$  CaO as



Fig. 2. Effect of specific surface area on the desulfurization activity of fly ash/CaO/CaSO<sub>4</sub> sorbents. Feed gas composition at 100 °C: 1000 ppm SO<sub>2</sub>, 500 ppm NO, 5%  $O_2$ , 12%  $CO_2$ , and the balance N<sub>2</sub>. The solid line is calculated from Eq. (1).

the specific surface area was increased from 22.5 to 133.3 m<sup>2</sup>/g. Within this specific surface area range, the exponential trend can be described quantitatively by the following empirical equation with a correlation coefficient of 0.991:

Desulfurization activity

$$
= 36.4 \exp(0.014 \times \text{specific surface area})
$$
 (1)

Although Eq. (1) is useful for predicting the effect of specific surface area on sorbent desulfurization activity, it is obvious that it should not be used for extrapolation purposes due to the empirical nature of the expression. This observation corroborates the finding of Lin et al. [\[11\]](#page-6-0) which showed that  $SO_2$  capture by sorbents prepared from fly ash/Ca(OH)<sub>2</sub> increased with increasing specific surface area. However, they found a linear correlation between desulfurization activity and specific surface area. It may be concluded that specific surface area plays a key role in enhancing the desulfurization capacity of sorbents derived from either fly ash/CaO/CaSO4 or fly ash/Ca(OH) $_2$ . Having established the influence of specific surface area on  $SO_2$  capture, a reference sorbent prepared from fly ash/CaO/CaSO<sub>4</sub> with a specific surface area of 64.5 m<sup>2</sup>/g was used in all desulfurization experiments reported below.

# *3.2. Comparison of fly ash/CaO/CaSO4 sorbent and starting materials*

 $SO<sub>2</sub>$  removal by the reference sorbent and the three starting materials is shown in [Fig. 3. T](#page-3-0)he fixed bed reactor was subjected to a feed gas consisting of 1000 ppm  $SO_2$ , 500 ppm NO, 5%  $O_2$ , 12% CO<sub>2</sub>, and the balance N<sub>2</sub> at 150 °C. As can be seen in [Fig. 3,](#page-3-0) the reference sorbent easily outperformed its base components in removing  $SO_2$ . The sorbent removed 100% of the  $SO_2$  in the feed gas during the first 30 min of reaction. The  $SO_2$  started to breakthrough at reaction time > 30 min, resulting in a gradual decrease in the percent removal of  $SO<sub>2</sub>$  with increasing reaction time. The percent removal of  $SO<sub>2</sub>$  eventually reached zero after 150 min of reaction. In contrast, each of the three starting materials; coal fly ash, CaO, and CaSO4, did not exhibit any significant desulfurization activity. These results most likely reflect the fact that the specific surface area of the reference sorbent (64.5 m<sup>2</sup>/g) was much larger than those of the starting materials (Table 1,  $1.5-5.6$  m<sup>2</sup>/g).

<span id="page-3-0"></span>

Fig. 3. Comparison of the desulfurization activity of the reference sorbent and its base components. Feed gas composition at  $150^{\circ}$ C:  $1000$  ppm SO<sub>2</sub>, 500 ppm NO, 5%  $O_2$ , 12%  $CO_2$ , and the balance  $N_2$ .

The surface morphology of the reference sorbent before and after SO<sub>2</sub> capture was studied through SEM. As shown in Fig. 4a, the hydration process produced a sorbent with a highly porous surface. After reacting with  $SO<sub>2</sub>$ , Fig. 4b shows that the porous surface of the sorbent was no longer discernible. The pore volume distributions of the reference sorbent before and after desulfurization are shown in Fig. 5. For the fresh sorbent, mesopores with an average pore size of 41.8 nm appeared to be the major contributor to the total pore volume. After desulfurization, a large portion of the pore volume had disappeared and the average pore size dropped to 28.6 nm, indicating significant pore filling by the reaction product. It may be concluded that the surface of the spent sorbent was covered by a layer of the reaction product CaSO4, reducing its porosity.



Fig. 5. Pore volume distributions of the reference sorbent before and after  $SO_2$ capture.

The chemical composition of the reference sorbent before and after desulfurization was studied through XRD analysis. The XRD spectra are shown in Fig. 6. The peaks in the spectra can be assigned as follows: peaks  $a =$  calcium aluminate silicate hydrate compound  $(Ca_2(Si_9A1_3)O_{24} \cdot 8H_2O)$ , peaks b = calcium carbonate (CaCO<sub>3</sub>), and peaks d = calcium sulfate (CaSO<sub>4</sub>).  $Ca_2(Si_9Al_3)O_{24}\cdot 8H_2O$  and  $CaCO_3$  are the products of the hydration process while the CaSO4 detected in the fresh sorbent is the unreacted base material. The absence of silica  $(SiO<sub>2</sub>)$  and alu- $\min(A_1A_2O_3)$  in Fig. 6a, the two main components of coal fly ash, suggests that they have reacted completely with CaO/CaSO4 to form  $Ca_2(Si_9Al_3)O_{24}\cdot 8H_2O$  which reacts with  $SO_2$ . After desulfurization, Fig. 6b shows noticeable changes in the spectrum of the sorbent. It can be seen that peaks a and b became weaker, indicating that Ca<sub>2</sub>(Si<sub>9</sub>Al<sub>3</sub>)O<sub>24</sub>.8H<sub>2</sub>O and CaCO<sub>3</sub> were converted to other chemical forms. Since the intensity of peaks d for CaSO4



Fig. 4. SEM micrographs of the reference sorbent (a) before and (b) after  $SO_2$ capture.



Fig. 6. XRD spectra of the reference sorbent (a) before and (b) after  $SO_2$  capture.

<span id="page-4-0"></span>

Fig. 7. Effect of reaction temperature on the percent removal of  $SO<sub>2</sub>$  by the reference sorbent. Feed gas composition:  $1000$  ppm  $SO_2$ ,  $500$  ppm NO,  $5\%$   $O_2$ , 12%  $CO<sub>2</sub>$ , and the balance  $N<sub>2</sub>$ .

increased, it may be concluded that the two compounds reacted with  $SO<sub>2</sub>$  to produce  $CaSO<sub>4</sub>$ . As already mentioned, the formation of CaSO4 led to a loss of sorbent porosity through CaSO4 deposit in the sorbent pores, as can be seen in the SEM image of [Fig. 4b.](#page-3-0)

## *3.3. Effect of reaction temperature*

Fixed bed experiments were performed to examine the effects of reaction temperature,  $SO<sub>2</sub>$  concentration, and NO concentration on the desulfurization activity of the reference sorbent. Fig. 7 shows the effect of reaction temperature (60–300  $\degree$ C) on the percent removal of  $SO_2$  from a feed gas containing 1000 ppm SO2. As can be seen in Fig. 7, the desulfurization activity increased when the temperature was raised from 60 to 200 ◦C but further increases in temperature beyond 200 ◦C did not seem to enhance the desulfurization activity. A possible explanation for this observation is that the hydrated water contained within the sorbent was released upon exposure to the hot gas stream, creating new pores which increased the surface area of the sorbent [\[15\].](#page-6-0) As already mentioned, desulfurization activity increases with increasing specific surface area.

When the temperature was raised by  $40^{\circ}$ C from 60 to 100  $^{\circ}$ C, the breakthrough time increased from 12 to 16 min, a rather modest increase in the sorbent's ability to maintain  $100\%$  SO<sub>2</sub> removal. This observation suggests that not much of the hydrated water trapped within the sorbent could be released at temperatures below the boiling point of water to create additional surface area. The weak effect of reaction temperature below  $100\degree\text{C}$ on the desulfurization activity of fly  $ash/Ca(OH)_2$  sorbent has also been reported by Liu et al. [\[16\].](#page-6-0) By contrast, raising the temperature by  $50^{\circ}$ C from 100 to  $150^{\circ}$ C increased the breakthrough time from 16 to about 30 min; the sorbent could maintain 100% removal of  $SO<sub>2</sub>$  for a much longer period of time. Similar enhancement in desulfurization activity was observed when the temperature was raised from 150 to 200 ◦C. This implies that the temperature range of 100–200 ◦C was able to cause significant release of the hydrated water of the sorbent which led to creation of additional pore volume and hence specific surface area for desulfurization. However, the breakthrough time remained somewhat constant when the temperature was increased from 200 to 300 °C. Two possible reasons may account for this observation. First, at temperatures >  $200^{\circ}$ C, surface coverage of the



Fig. 8. Effect of  $SO_2$  concentration on the desulfurization activity of the reference sorbent. Feed gas composition at  $100\degree C$ : 500–2500 ppm SO<sub>2</sub>, 500 ppm NO, 5%  $O_2$ , 12%  $CO_2$ , and the balance  $N_2$ .

sorbent by the resulting reaction product  $CaSO<sub>4</sub>$  prevented further release of hydrated water. Second, all the hydrated water of the sorbent had already been released when the temperature reached 200 ◦C.

## *3.4. Effect of SO2 concentration*

The effect of  $SO<sub>2</sub>$  concentration on the desulfurization activity of the reference sorbent is shown in Fig. 8. The desulfurization activity decreased linearly when the  $SO<sub>2</sub>$  concentration was increased from 500 to 2500 ppm. The slope of the line in Fig. 8 estimated by a linear regression with a correlation coefficient of 0.979 was  $-0.052 \text{ min/(g CaO ppm SO}_2)$ . The drop in the desulfurization activity is a direct consequence of exposing a fixed amount of sorbent to increasing  $SO_2$  concentration in the feed gas.

## *3.5. Effect of NO concentration*

Fig. 9 shows the effect of NO concentration on the extent of SO2 removal by the reference sorbent. The presence of NO in the feed gas had a beneficial effect on the desulfurization activity. In the absence of NO the time that the sorbent could maintain 100% removal of  $SO<sub>2</sub>$  was only about 9 min; however, when 250 ppm NO was included in the feed gas the breakthrough time increased to about 20 min. Further enhancement of the breakthrough time was observed using a feed gas containing 500 ppm NO. It has been postulated that the presence of NO in the feed



Fig. 9. Effect of NO concentration on the percent removal of  $SO_2$  by the reference sorbent. Feed gas composition at 100 °C: 1000 ppm SO<sub>2</sub>, 0–750 ppm NO, 5%  $O_2$ , 12%  $CO_2$ , and the balance  $N_2$ .

<span id="page-5-0"></span>

Fig. 10. FTIR spectra of spent sorbent (a) in the absence of NO and (b) in the presence of 500 ppm NO.

gas resulted in the formation of  $NO<sub>2</sub>$  which oxidizes  $SO<sub>2</sub>$  to  $SO<sub>3</sub>$ . The resulting  $SO<sub>3</sub>$  can react with calcium compounds in the sorbent several times faster than  $SO<sub>2</sub>$  to form sulfate salts. However, an unexpected phenomenon occurred when the feed NO concentration was increased to 750 ppm. As can be seen in [Fig. 9,](#page-4-0) the breakthrough behavior at 750 ppm NO was similar to that at 500 ppm NO up to a reaction time of 70 min. After 70 min, the percent removal of  $SO<sub>2</sub>$  began to increase with time, reaching a maximum before declining slowly. This unexpected behavior has also been observed in a previous study using slaked lime sorbent and a feed gas containing  $800$  ppm  $SO<sub>2</sub>$  and 800 ppm NO [\[17\].](#page-6-0) The reason for this behavior is not known and further research is needed in order to shed light on this phenomenon.

The role of NO in  $SO_2$  removal was studied through FTIR analysis. The FTIR spectra of spent sorbent in the absence of NO and in the presence of 500 ppm NO are shown in Fig. 10. In the absence of NO, Fig. 10a shows that the vibration absorption peaks ascribed to SO<sub>3</sub><sup>2–</sup> (sulfite) and SO<sub>4</sub><sup>2–</sup> (sulfate) were detected in the spent sorbent. The peak for  $SO_3^2$ <sup>-</sup> is probably

due to the absorption of  $SO<sub>2</sub>$  in the form of sulfite salts, while the  $SO_4^2$ <sup>-</sup> peak is due to the unreacted base material used in synthesizing the sorbent or it could also be due to the absorption of  $SO<sub>2</sub>$  in the form of sulfate salts. In the presence of 500 ppm NO, the vibration absorption peaks ascribed to  $SO_3^2$ <sup>-</sup> and  $SO_4^2$ <sup>-</sup> were still detected in the spent sorbent, as shown in Fig. 10b. Comparison of the two spectra indicates that the peak for  $SO_4^2$ <sup>-</sup> in Fig. 10b became more intense while the peak for  $SO_3^2$ <sup>-</sup> was weaker. The increase in the intensity of the  $SO_4^2$  peak and decrease in the intensity of the  $SO_3^2$ <sup>-</sup> peak may be attributed to the oxidizing effect of NO. Most of the  $SO<sub>2</sub>$  was absorbed in the form of sulfate salts instead of sulfite salts in the presence of NO. Further, a weak vibration absorption peak ascribed to  $NO<sub>3</sub><sup>-</sup>$  (nitrate) was also detected, as shown in Fig. 10b, indicating that a small amount of NO was absorbed in the form of nitrate salts.

#### **4. Conclusions**

The desulfurization activity of sorbents synthesized from coal  $fly$  ash/CaO/CaSO<sub>4</sub> was observed to increase exponentially with increasing specific surface area. The extent of  $SO<sub>2</sub>$  capture by a reference sorbent with a specific surface area of  $64.5 \text{ m}^2/\text{g}$  was much higher than those of its starting materials. Enhancement of desulfurization activity by reaction temperature and inclusion of NO in the feed gas was also observed. SEM showed that  $SO<sub>2</sub>$  capture resulted in a reduction of sorbent porosity through surface coverage by the reaction product CaSO4. XRD and FTIR analyses confirmed that  $SO<sub>2</sub>$  was absorbed in the form of sulfate salts.

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#### **References**

- [1] N. Karatepe, A.E. Mericboyu, S. Kucukbayrak, Preparation of fly ash  $Ca(OH)_2$  sorbents by pressure hydration for  $SO_2$  removal, Energy Sources 20 (1998) 945–953.
- [2] N. Karatepe, A. Ersoy-Mericboyu, U. Demirler, S. Kucukbayrak, Determination of the reactivity of  $Ca(OH)_2$  fly ash sorbents for  $SO_2$  removal from flue gases, Thermochim. Acta 319 (1998) 171–176.
- [3] N. Karatepe, A. Ersoy-Mericboyu, S. Kucukbayrak, Activation of Ca(OH)2 using different siliceous materials, Environ. Technol. 20 (1999) 377–385.
- [4] H. Tsuchiai, T. Ishizuka, T. Ueno, H. Hattori, H. Kita, Highly active absorbent for SO<sub>2</sub> removal prepared from coal fly ash, Ind. Eng. Chem. Res. 34 (1995) 1404–1411.
- [5] T. Ishizuka, H. Tsuchiai, T. Murayama, T. Tanaka, H. Hattori, Preparation of active absorbent for dry-type flue gas desulfurization from calcium oxide, coal fly ash, and gypsum, Ind. Eng. Chem. Res. 39 (2000) 1390–1396.
- <span id="page-6-0"></span>[6] J. Fernandez, J. Renedo, A. Garea, J. Viguri, J.A. Irabien, Preparation and characterization of fly ash/hydrated lime sorbents for  $SO_2$  removal, Powder Technol. 94 (1997) 133–139.
- [7] A. Garea, I. Fernandez, J.R. Viguri, M.I. Ortiz, J. Fernandez, M.J. Renedo, J.A. Irabien, Fly-ash/calcium hydroxide mixtures for SO<sub>2</sub> removal: structural properties and maximum yield, Chem. Eng. J. 66 (1997) 171–179.
- [8] M.J. Renedo, J. Fernandez, A. Garea, A. Ayerbe, J.A. Irabien, Microstructural changes in the desulfurization reaction at low temperature, Ind. Eng. Chem. Res. 38 (1999) 1384–1390.
- [9] M.J. Renedo, J. Fernandez, Preparation, characterization, and calcium utilization of fly ash/Ca(OH)<sub>2</sub> sorbents for dry desulfurization at low temperature, Ind. Eng. Chem. Res. 41 (2002) 2412–2417.
- [10] R.-B. Lin, S.-M. Shih, C.-F. Liu, Characteristics and reactivities of  $Ca(OH)_2$ /silica fume sorbents for low-temperature flue gas desulfurization, Chem. Eng. Sci. 58 (2003) 3659–3668.
- [11] R.-B. Lin, S.-M. Shih, C.-F. Liu, Structural properties and reactivities of  $Ca(OH)_2/fly$  ash sorbents for flue gas desulfurization, Ind. Eng. Chem. Res. 42 (2003) 1350–1356.
- [12] T. Ishizuka, T. Yamamoto, T. Murayama, T. Tanaka, H. Hattori, Effect of calcium sulfate addition on the activity of the absorbent for dry flue gas desulfurization, Energy Fuels 15 (2001) 438– 443.
- [13] J. Fernandez, M.J. Renedo, A. Pasquera, J.A. Irabien, Effect of CaSO<sub>4</sub> on the structure and use of  $Ca(OH)/fly$  ash sorbents for  $SO_2$  removal, Powder Technol. 139 (2001) 201–205.
- [14] J. Fernandez, M.J. Renedo, Study of the influence of calcium sulfate on fly ash/Ca(OH)<sub>2</sub> sorbents for flue gas desulfurization, Energy Fuels 17 (2003) 1330–1337.
- [15] W. Jozewicz, J.C.S. Chang, T.G. Brna, C.B. Sedman, Reactivation of solids from furnace injection of limestone for sulfur dioxide control, Environ. Sci. Technol. 21 (1987) 664–670.
- [16] C.-F. Liu, S.-M. Shih, R.-B. Lin, Kinetics of the reaction of  $Ca(OH)_2/fly$ ash sorbent with  $SO_2$  at low temperatures, Chem. Eng. Sci. 57 (2002) 93–104.
- [17] M. Sakai, C. Su, E. Sasaoka, Simultaneous removal of SO*<sup>x</sup>* and NO*<sup>x</sup>* using slaked lime at low temperature, Ind. Eng. Chem. Res. 41 (2002) 5029–5033.